Contribution of particulate nitrate to airborne measurements of total reactive nitrogen

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[1] Simultaneous measurements of speciated, total reactive nitrogen (NO_v) and particulate NO₃ (particle diameter <1.3 μm) were made on board the NASA P-3B aircraft over the western Pacific in February-April 2001 during the Transport and Chemical Evolution over the Pacific (TRACE-P) experiment. Gas-phase and particulate NO_v was measured using a gold tube catalytic converter. For the interpretation of particulate NO₁, conversion efficiencies of particulate NH₄NO₃, KNO₃, NaNO₃, and Ca(NO₃)₂ were measured in the laboratory. Only NH₄NO₃ showed quantitative conversion, and its conversion efficiency was as high as that for HNO₃. NO_v measured on board the aircraft was found to be systematically higher by 10-30% than the sum of the individual NO_v gas components $(\sum (NO_v)_i)$ at 0–4 km. Particulate NO_3^- concentrations measured by a particle-into-liquid sampler (PILS) were nearly equal to $NO_v - \sum (NO_v)_i$ under low-dust-loading conditions. The PILS data showed that the majority of the particulate NO₃ was in the form of NH₄NO₃ under these conditions, suggesting that NH₄NO₃ particles were quantitatively converted to detectable NO by the NO_v converter, consistent with the laboratory experiments. The contribution of particulate NO₃ to NO_v was most important at 0-2 km, where NO_3^- constituted 10–30% of NO_v during TRACE-P. On average, the amounts of particulate NO₃ and gas-phase HNO₃ were comparable in this region.

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Introduction

[2] Reactive nitrogen species play a central role in determining the levels of ozone and hydroxyl radicals in the troposphere [e.g., Crutzen, 1979; Liu et al., 1987; Chameides et al., 1992]. Total reactive nitrogen (NO_v) in the troposphere is generally composed of NO, NO₂, peroxyacetyl nitrates (PANs), nitric acid (HNO₃), HONO,

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HO₂NO₂, NO₃, N₂O₅, organic nitrate, and particulate nitrate (NO₃⁻). One of the major uncertainties in evaluating the budget of reactive nitrogen is due to a lack of fast response and accurate measurements of both gas-phase HNO₃ and particulate NO₃. This point is also important in evaluating loss rates of NO_v, because the deposition velocity for particulate NO_3^- with particle diameters $(D_p) < 1 \mu m$ is 10 times smaller than that of gas-phase HNO₃ [Warneck, 1988].

[3] Reaction of NH₃ with HNO₃ produces ammonium nitrate (NH₄NO₃) particles generally with $D_p < 1 \mu m$. The thermodynamic equilibrium between HNO₃ and NH₄NO₃ shows a strong dependence on ambient temperature and relative humidity characteristic of the troposphere. Particulate NO_3^- with $D_p > 1 \mu m$ can be produced by adsorption of HNO₃ on basic soil [Wolff, 1984] or sea-salt particles [Savoie and Prospero, 1982]. In the upper troposphere, HNO₃ uptake on cirrus ice cloud particles followed by gravitational settling may reduce the abundance of HNO₃ [Lawrence and Crutzen, 1998]. At high latitudes in the lower stratosphere, HNO₃ is irreversibly removed by sedimentation of large HNO₃-containing polar stratospheric cloud (PSC) particles (denitrification) formed at very low temperatures [Solomon et al., 1986]. Therefore it is important to assess the contribution of particulate NO_3^- to NO_v in order to understand distributions of reactive nitrogen.

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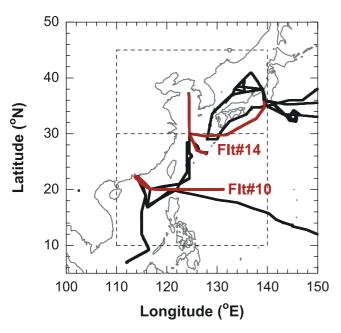


Figure 1. Flight tracks near the Asian continent during TRACE-P and regions where flights 10 and 14 were made (see text for details).

- [4] To date, a number of attempts have been made to quantify the budget of NO_v over a variety of regions. Catalytic conversion of NO_v compounds on the surface of a heated gold tube, followed by chemiluminescence detection of NO [Fahev et al., 1985] has been widely used for airborne measurements, because of its fast response and high sensitivity [e.g., Hübler et al., 1992; Ridley et al., 1994; Sandholm et al., 1994; Kondo et al., 1997a]. Over the western Pacific, NO_x, PAN, and HNO₃ obtained by airborne measurements together accounted for 90% of concurrently measured NO_v up to 7 km [Kondo et al., 1997b]. Measurements of NO_v and the sum of reactive nitrogen species in the upper troposphere over the North Atlantic [Talbot et al., 1999] showed that 90% of the measured NO_v was accounted for by measurements of the individual NO_v species.
- [5] Over the North American continent in summer, aircraft measurements showed that the individually measured species ($NO_x + HNO_3 + PAN_5 + C_1-C_5$ alkyl nitrate) constituted more than 90% of the measured NO_y on average at altitudes up to 7 km [Neuman et al., 2002; Ryerson et al., 2003; Parrish et al., 2004]. They concluded that the sum of the individual NO_y constituents agreed with the measured NO_y within the instrument uncertainties over the region. Day et al. [2003] recently showed that the sum of total alkyl nitrates (RONO₂) routinely accounted for 10-20% of NO_y , using annual ground-based measurements at a rural site in California
- [6] In situ aircraft measurements of NO_y have been made through forward and/or rearward facing inlets. NO_y measured through forward facing inlets represents amplified particulate NO_y plus gas-phase NO_y. Fahey et al. [2001] observed large, HNO₃-containing particles with D_p = 10–20 μ m in the winter Arctic lower stratosphere by in situ aircraft measurements of NO_y. Similarly, NO_y contained in cirrus cloud ice particles has been measured in the upper

troposphere [Weinheimer et al., 1998; Kondo et al., 2003] and in the lowermost stratosphere [Feigl et al., 1999]. However, the response of NO_y measurement techniques or their sampling systems to particulate NO_3^- in the lower troposphere is not fully understood. In particular, a quantitative assessment of the contribution of particulate NO_3^- to NO_y sampled through rearward facing inlets has not been made on airborne platforms in the lower troposphere. The major uncertainties are (1) particle collection efficiency of the rearward facing inlet and (2) conversion efficiency of particulate NO_3^- in the NO_y converter.

[7] In this paper, we present evidence of the detection of particulate NO_3 by an airborne NO_y instrument during the NASA Transport and Chemical Evolution over the Pacific (TRACE-P) experiment. We focus on evaluating the contribution of NO_3 to NO_y measurements over the western Pacific.

2. Experiment

[8] During the TRACE-P experiment (February-April 2001), NO_y and individual NO_y compounds were measured on board the P-3B aircraft over the Pacific at 0-7 km [Kondo et al., 2004]. We use the data obtained in the region of $10^\circ-45^\circ N$, $110^\circ-140^\circ E$, where most of the P-3B sampling was made (Figure 1). The data used in this study were obtained during flights 10-20 (9 March to 4 April 2001). Since details of each measurement technique are summarized elsewhere [see *Miyazaki et al.*, 2003, Table 1], we describe here the NO_y and aerosol measurements.

2.1. Gas-Phase and Particulate NO_v

[9] NO_v was measured using an NO-O₃ chemiluminescence technique combined with a gold tube heated to 300°C with addition of CO, catalytically converting NO_v compounds into NO [Kondo et al., 1997a]. The NO_v converter unit was mounted on the side of the window plate inside the aircraft cabin to minimize the length of the inlet tubes. During in-flight calibrations, known concentrations of NO and NO₂ in N₂ were added into the sample air to calibrate the absolute sensitivity and conversion efficiency. The estimated uncertainty of the NO mixing ratios of the calibration gas was 2%, and the uncertainty of the flow rates was estimated to be $\pm 2\%$. The error from the artifact for NO_{ν} is estimated to be $\pm 2\%$ for NO_{ν} mixing ratios of 500 parts per trillion by volume (pptv). The HNO₃ conversion efficiency was higher than 95%, which was determined by both laboratory tests and in-flight calibration during the previous aircraft missions [Kondo et al., 1997a; Koike et al., 2000]. It was measured to be stable over time within $\pm 5\%$ [Kondo et al., 1997a]. The conversion efficiencies of HCN and CH₃CN were 2% and 1-2%, respectively [Koike et al., 2000]. From the observed values of HCN (<220 pptv) and CH₃CN (<150 pptv) at 0–6 km during TRACE-P [Singh et al., 2003], the uncertainty associated with the conversion of these species is estimated to be less than 2%. Overall, the accuracy of the NO_v measurement was estimated to be

[10] In this study, ΔNO_y is defined as the difference between the measured NO_y and $\sum (NO_y)_i$,

$$\Delta NO_y = NO_y - \sum (NO_y)_i \tag{1}$$

Table 1. Approximate Stokes Numbers Calculated for the P-3B NO_{ν} Inlet at a Pressure Altitude of 950 hPa

	Value
$D_p = 0.3 \mu m$	0.01
$D_{p} = 0.5 \ \mu m$	0.03
$D_{p} = 1.0 \ \mu m$	0.09
$\underline{D_p} = 1.5 \ \mu m$	0.19

where $\sum (NO_y)_i$ is defined as the sum of individually measured gas-phase NO_y compounds.

$$\sum (NO_y)_i = NO + NO_2 + HNO_3 + PAN + PPN + C_1 - C_5 \text{ alkyl nitrates}$$
(2)

The uncertainties of $\sum (NO_y)_i$ and ΔNO_y were estimated to be $\pm 14\%$ and $\pm 32\%$, respectively, by combining possible errors of the individual instruments.

[11] During TRACE-P, two NO_v inlets were located on the right window plate of the cabin, one facing rearward and the other facing forward. The inlet of the NO_v instrument was made of 3/8-inch outer diameter Teflon tubing heated to 50°C for rearward and 100°C for forward upstream of the NO_v converter unit. The tubing from the inlet to the gold tube converter is not straight but is curved (angle 90 deg, radius 5 cm). In order to estimate the collection efficiency of particles for the rearward NO_v inlet, we made a simple calculation based on the Stokes number, which is the ratio of a particle's "persistence" to the size of the obstacle, often used to characterize inertial impaction. The rearward facing inlet allows the collection of particles with Stokes numbers less than about 0.2, which track streamlines of sample air because of their negligibly small inertia. At Stokes numbers greater than 0.2, particles do not follow the streamlines that curve into the rearward facing inlet. Approximate Stokes numbers were estimated for the P-3B NO_v sampling inlet by a simple aerodynamical calculation (Table 1). On the basis of these calculations, we estimated that the NO_v inlet discriminates against particles with D_p larger than 1.5 and 1.1 µm at altitudes of 0.5 and 6 km, respectively. The collection efficiency of particles with $D_p > 1.5 \mu m$ is estimated to be almost zero. It should be noted that effects of turbulence were not taken into account in the above

estimate. Turbulence introduces uncertainties in the estimate of the cutoff size, although they are not quantified in this work.

[12] During TRACE-P, NO_y was also sampled through a forward facing inlet using an independent detection channel. The forward facing inlet was subisokinetic and collected particles including those with $D_p > 1.5 \mu m$ that were unlikely sampled by the rearward NO_y inlet. The NO_y measured through the forward facing inlet $(NO_y$ (f)) represents the sum of gas-phase NO_y and amplified particulate NO_3^- . When particulate NO_3^- is absent, NO_y signals from NO_y (f) and NO_y are identical. When encountered, particulate NO_3^- including larger particles $(D_p > 1.5 \mu m)$, NO_3^- (L), can be defined as follows:

$$NO_3^-(L) = \left(NO_y(f) - \sum (NO_y)_i\right) / EF$$
 (3)

EF represents the enhancement factor for the forward facing inlet, which depends on the size of the particles due to subisokinetic sampling [Fahey et al., 1989].

2.2. Laboratory Experiments

[13] In order to estimate the conversion efficiency of particulate NO₃⁻ in the gold tube converter, we conducted laboratory experiments using an aerosol generation system (Figure 2). Because particulate NO₃ is observed typically in the form of NH₄NO₃, KNO₃, NaNO₃, and Ca(NO₃)₂ in the troposphere, we measured the conversion efficiencies of the particles in these forms in the laboratory. Dry particles made of these chemical compositions were generated using an atomizer (TSI model 3076) and diffusion dryer. A differential mobility analyzer (DMA) (TSI model 3080) was used to produce monodisperse particles with a size range between 300 and 400 nm. The number concentrations were monitored by a condensation particle counter (CPC) (TSI model 3022A). Mass concentrations of nitrate particles were calculated using the density (p), diameter, number concentrations, and shape factor [Jayne et al., 2000; DeCarlo et al., 2004]. The results are summarized in Table 2. The conversion efficiency for NH₄NO₃ was 91 \pm 9% at NO₃ mixing ratios of 2000-5000 pptv. The derived conversion efficiency for particulate NH₄NO₃ was as high as that for HNO₃. The difference between the two conversion efficiencies is to within the measurement uncertainties. Uncertainties of the

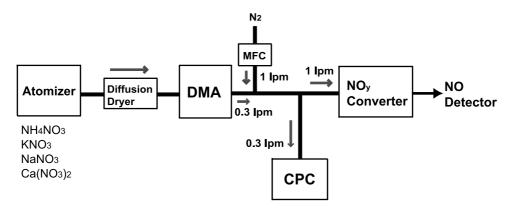


Figure 2. Schematic of the laboratory setup to generate and measure particulate NO₃. The number concentration of generated aerosol was measured by a condensation particle counter (CPC).

Table 2. Conversion Efficiencies of Particulate NO₃ Species Measured in the Laboratory^a

	NH ₄ NO ₃	KNO ₃	NaNO ₃	Ca(NO ₃) ₂
NO ₃ Conversion	90.7 ± 9.0	1.2 ± 0.5	2.0 ± 0.3	3.4 ± 0.8
Efficiency, % Assumed Density ρ, g/cm ³	1.72	2.11	2.66	2.36
Boiling point/Melting point, °C	210 (b)	334 (m)	380 (b)	561 (m)

^aThe assumed density and boiling/melting points for different forms of particles are also shown. Here, "(b)" and "(m)" denote boiling point and melting point, respectively.

NH₄NO₃ conversion efficiency due to evaporation of NH₄NO₃ particles in the inlet system are small. The residence time of particulate NO₃ between the atomizer and NO_v converter is about 1 s, which is much shorter than the typical timescale of a few minutes required for evaporation of particulate NO₃ [e.g., Neuman et al., 2003]. In fact, our recent ground-based measurements showed that the ambient HNO₃ concentrations were smaller than 8% of the particulate NO_3^- levels (~ 1 ppbv) during nighttime, when the ambient temperature was about 15°C (K. Kita et al., A chemical ionization mass spectrometer for ground-based measurement of nitric acid, submitted to Journal of Atmospheric and Oceanic Technology, 2005), which is similar to the room air temperature during the laboratory experiments. By contrast, the conversion efficiencies for KNO₃, NaNO₃, and Ca(NO₃)₂ were as low as $1.2 \pm 0.5\%$, $2.0 \pm 0.3\%$, and $3.4 \pm 0.8\%$, respectively.

[14] These results are consistent with the boiling or melting points of the particulate NO₃⁻ species (Table 2). NH₄NO₃ (boiling point 210°C) should be vaporized at the temperature of the NO_y converter (300°C). On the other hand, the melting or boiling points of KNO₃, NaNO₃, and (Ca(NO₃)₂) are higher than the temperature of the NO_y converter, leading to significantly low conversion efficiencies.

2.3. Aerosol Measurements

[15] Aerosol chemical composition (NO_3^- , SO_4^{2-} , Cl^- , NH_4^+ , Na^+ , K^+ , and Ca^{2+}) was measured every 4 min, using a particle-into-liquid sampler coupled with ion chromatography analysis (PILS-IC) [Weber et al., 2001; Orsini et al., 2003]. The PILS employs a steam saturator to grow the aerosol to sizes that result in liquid impaction followed by measurement of the major ions using IC. For the PILS, the collection efficiency for $D_p < 0.7 \mu m$ was estimated to be 90%, decreasing to \sim 50% for $D_p = 1.3 \mu m$ (upper size cut), which was determined by laboratory calibration [Ma et al., 2004]. The efficiency for $D_p > 3 \mu m$ was nearly zero. The uncertainty of the NO₃ measurement was estimated to be $\pm 15\%$. The P-3B used a shrouded inlet for the aerosol measurements, which maintained an attack angle independent of turbulence characteristics in the diffuser cone [Clarke et al., 2004]. The aerosol inlet was kept isokinetic during flight by adjusting flow rates as flight parameters

[16] Upper size cuts of the NO_y and PILS systems are shown in Figure 3. Basically, the size cutoff of the PILS is similar to that of the NO_y instrument. However, the PILS measures particles within a size range of $D_p = 1.5-3 \mu m$,

although the collection efficiency within this range is less than 50%. On the other hand, it is unlikely that particles with $D_p = 1.5-3~\mu m$ were sampled by the NO_y instrument. We discuss the difference in the threshold of particle size in section 4.

17] Ambient aerosol size distributions were measured by a laser optical counter (OPC, PMS LAS-X, Boulder, CO) [Clarke et al., 2004]. The OPC provided size distributions of "dry" (heated to 40°C) particles between 0.1 and 20.0 μm. The terms "fine particle" and "coarse particle" in this paper are defined as particles within size ranges of $D_p < 1$ μm and $D_p > 1$ μm, respectively. Ambient wet particle size distributions are estimated by assuming that particles with $D_p < 0.6$ μm absorb water as a typical marine boundary layer aerosol [Swietlicki et al., 2000], and particles with $D_p > 0.6$ μm are treated as dust and do not absorb water [Weber et al., 2003]. All of the P-3B data were merged into the time interval of aerosol sampling (~4 min).

3. Particulate NO₃ Detected by the NO_y Instrument

[18] Figure 4 shows time series plots of NO_y , ΔNO_y , NO_3^- , fine-particle volume, NH_4^+ , and K^+ obtained on 9 March (flight 10) and 18 March (flight 14) 2001. The flight tracks are also shown in Figure 1. A large enhancement of NO_y up to 2500 pptv was observed at an altitude of 3 km during flight 10. The ΔNO_y and NO_3^- concentrations reached their maximum values in the same air masses. An interesting feature is that K^+ also showed its maximum (\sim 900 pptv) in these plumes, which were identified as originating from biomass burning in southeast Asia by using back trajectories and satellite fire maps [$Ma\ et\ al.$, 2003; $Kondo\ et\ al.$, 2004]. $Ma\ et\ al.$ [2003] used observed K^+ /sulfate ratios to infer that 30% of total aerosol mass in Asian pollution plumes sampled

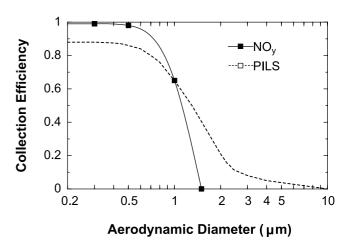


Figure 3. Upper size limits of particle measurements for the NO_y and PILS instruments. The NO_y transmission curve was derived on the basis of the calculated Stokes number assuming no turbulence. Turbulence introduces uncertainties in the estimate of the cutoff size, although they are not quantified in this work. The transmission curve for the PILS was determined experimentally by generating monodisperse calibration particles using a nebulizer/DMA and vibrating-orifice aerosol generator in the laboratory [*Ma et al.*, 2004].

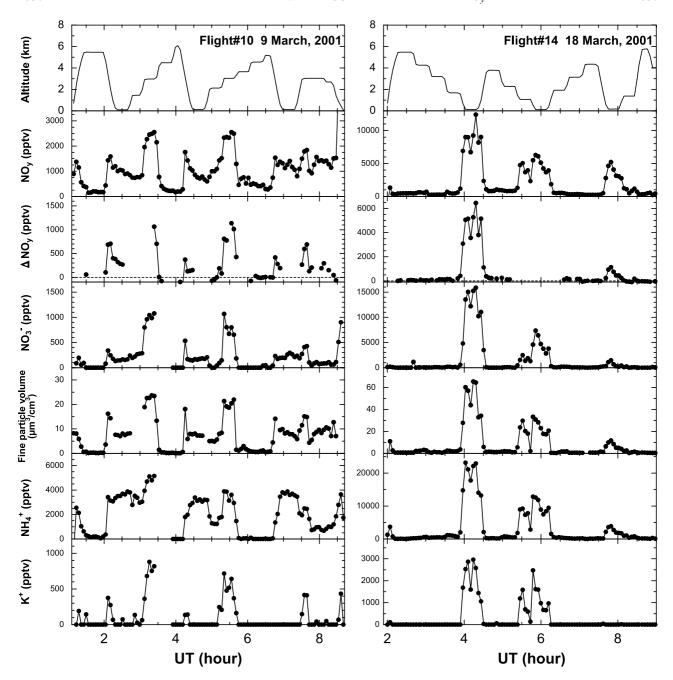


Figure 4. Time series plots for flight 10 and flight 14 with altitude, NO_{3} , ΔNO_{3} , fine-particle volume, NH_{4}^{+} , and K^{+} .

during TRACE-P originated from biomass or biofuel burning. During flight 14, very distinct plumes were observed and all species reached their highest concentrations at 0–1 km over the Yellow Sea (34°–36°N, see Figure 1). Peak concentrations for NO $_{\!\scriptscriptstyle J}$, Δ NO $_{\!\scriptscriptstyle J}$, and NO $_{\!\scriptscriptstyle 3}^-$ were 13,000, 6500, and 16,000 pptv, respectively. For both cases, fine-particle volumes were also enhanced (25–70 $\mu m/cm^3$) in these plumes.

[19] In order to evaluate the contribution of NO_3^- to ΔNO_y , their correlations at 0–4 km are shown in Figure 5. A majority of the NO_3^- and ΔNO_y (<2500 pptv) data points in Figure 5a lie along the 1:1 line, within the error bars, indicating that NO_3^- measured by the PILS-IC can account for the ΔNO_y . In Figure 5b, data points with ΔNO_y^- >

2500 pptv are also shown according to Ca^{2+} concentration. Ca^{2+} is generally considered to be derived from soil and crustal material and thus can be used as a tracer for dust. For the data with $\Delta NO_y < 2500$ pptv, which accounted for 97% of all the data points, dust loadings were low ([Ca^{2+}] < 200 pptv). On the other hand, NO_3^- with $\Delta NO_y > 2500$ pptv far exceeded ΔNO_y where dust loadings were high ([Ca^{2+}] > 400 pptv). These data with $\Delta NO_y > 2500$ pptv were all obtained over the Yellow Sea during flight 14 (Figure 4). A possible explanation of the excess NO_3^- relative to ΔNO_y is discussed in section 4.

[20] Figure 6 shows latitudinal distributions of (coarse-particle volume)/(fine-particle volume) ratios observed at 0-4 km during TRACE-P. Although variability was large,

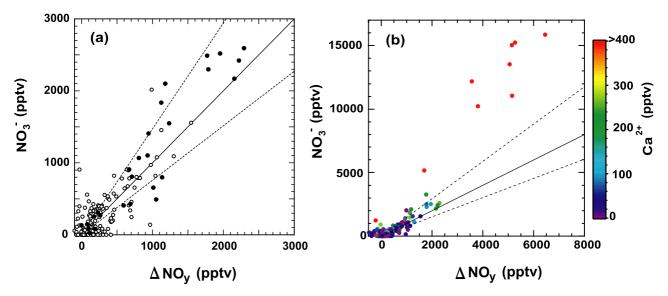


Figure 5. Linear relationship between particulate NO_3^- and ΔNO_y observed at 0–4 km for all flights during the TRACE-P period. Solid and dashed lines indicate the 1:1 line and the uncertainty of ΔNO_y , respectively. (a) Solid circles indicate the data in which $NO_3^- + SO_4^{2-}$ and NH_4^+ in equivalence balanced to within $\pm 30\%$. (b) The data points are color-coded according to the concentration of nss-Ca²⁺. Note that Figure 5a is a subset of the data points in Figure 5b and that the scales of the x and y axes in both figures are different.

the median ratio at $10^{\circ}-30^{\circ}N$ was lower than 1, suggesting that a majority of particles was in the fine mode ($D_p < 1 \mu m$), likely in biomass burning plumes [Kondo et al., 2004]. A dominance of fine particles in biomass burning plumes was previously observed, for instance, in Brazil [Reid and Hobbs, 1998]. By contrast, at 30°-45°N, coarse-particle volume was larger than fine-particle volume. Air masses with a large abundance of coarse particles observed in this region were heavily influenced by dust [Jordan et al., 2003a]. In fact, Jordan et al. [2003a] found that dust particles accounted for 77% of the total aerosol mass in the Chinese low-altitude outflow, typically observed at 30°-45°N during TRACE-P. Figure 7 presents a plot of the $NO_3^-/\Delta NO_v$ ratio versus the (coarse-particle volume)/ (fine-particle volume) ratio. At the $NO_3^-/\Delta NO_v$ ratio close to 1, fine-particle volume was generally larger than coarseparticle volume and dust loadings were low with [Ca²⁺] < 200 pptv. These results suggest that a majority of ΔNO_{ν} can be attributed to fine particles.

[21] Contributions of NO_y species not measured in this study to Δ NO_y are not significant in this study. Model calculations [Brasseur et al., 1999; Talbot et al., 1999] indicate that HO₂NO₂ is the most abundant unmeasured NO_y species. However, it is thermally stable only at the cold temperatures of the upper troposphere, and the HO₂NO₂/NO_y ratio is estimated to be lower than 0.01 below 4 km [Brasseur et al., 1999]. On the other hand, Δ NO_y is significantly large in the lower troposphere (0–4 km) in this case. Consequently, the contribution of HO₂NO₂ to Δ NO_y can be negligible in this study.

4. Chemical Form of Particulate NO₃

[22] In this section, we infer the chemical form of particulate NO_3^- detected by the NO_v instrument. Figure 8

shows average ionic composition at 0-4 km with ΔNO_y both less than and greater than 2500 pptv, where dust loadings were low with $[Ca^{2^+}] < 200$ pptv and high with $[Ca^{2^+}] > 400$ pptv, respectively. NH_4^+ was the main neutralizing agent of the acidic aerosol species, which accounted for 82 and 73% of the total cations with $\Delta NO_y < 2500$ pptv and $\Delta NO_y > 2500$ pptv, respectively. NH_4NO_3 forms by condensation of gas-phase HNO_3 with available NH_3 once all $SO_4^{2^-}$ is neutrized [Seinfeld and Pandis, 1998]. In order to examine this, the relationship between NO_3^- and NH_4^+ minus $SO_4^{2^-}$ in equivalence is shown in Figure 9. Overall, a

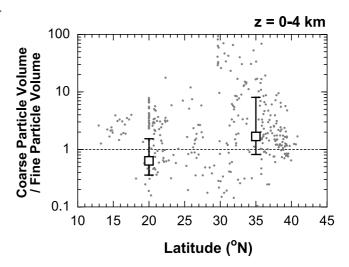
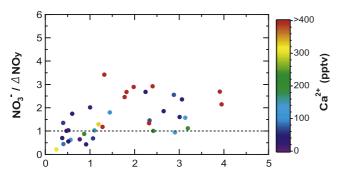


Figure 6. Latitudinal distributions of coarse ($D_p = 1-20~\mu m$)/fine ($D_p = 0.1-1~\mu m$) particle volume observed at 0-4 km. Median values are shown for $10^\circ-30^\circ N$ and $30^\circ-45^\circ N$. The data points with fine-particle volume <0.2 $\mu m^3/cm^3$ are not used.



Coarse particle volume / Fine particle volume

Figure 7. Relationship between $NO_3^-/\Delta NO_y$ and coarse $(D_p = 1-20 \mu m)/\text{fine}$ $(D_p = 0.1-1 \mu m)$ particle volume observed at 0-4 km. The data points are color-coded according to the concentration of nss-Ca²⁺. Dashed lines indicate that the $NO_3^-/\Delta NO_y$ ratio is 1. The data points with $\Delta NO_y < 100$ pptv and fine-particle volume $<0.2 \mu m^3/\text{cm}^3$ are not used.

positive correlation ($r^2 = 0.77$) with a slope of 1.10 shows that ion balance between NH₄⁺ minus SO₄²⁻ and NO₃⁻ in equivalence is nearly complete, within 10% on average, suggesting available NH₃ to form NH₄NO₃. For Δ NO_y < 2500 pptv, the slope was 0.97 with $r^2 = 0.66$ and 1.34 with $r^2 = 0.74$ for Δ NO_y > 2500 pptv. The data for which NO₃⁻ + SO₄²⁻ and NH₄⁺ in equivalence balanced within ±30% are plotted in Figure 5a (solid circles). The slope for the data points was 1.16 ($r^2 = 0.87$), lying close to the 1:1 line within the error bars, indicating that NH₄NO₃ was detected by the NO_y instrument.

[23] In order to estimate the relative contribution of cations other than NH_4^+ associated with NO_3^- , we consider the relative amounts of Ca^{2+} , K^+ , and Na^+ by using PILS chemical composition data. If we assume that all of the Ca^{2+} and K^+ take up SO_4^{2-} , measured NO_3^- is considered to be in the form of NH_4NO_3 , that is, NO_3^- as $NH_4NO_3 = NH_4^+ - SO_4^{2-}$ in equivalence (upper limit). By contrast, when all Ca^{2+} and K^+ are assumed to take up the measured NO_3^- , the

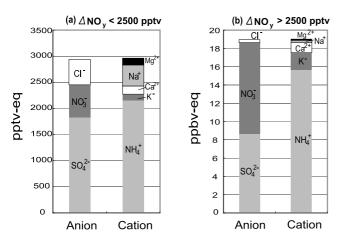


Figure 8. Ionic composition in pptv-equivalent with ΔNO_y (a) <2500 pptv and (b) >2500 pptv at 0–4 km. Note that the full scales of the y axes are different in Figures 8a and 8b.

amount of NO_3^- as NH_4NO_3 can be derived from $NO_3^ (Ca^{2^+} + K^+)$ in equivalence (lower limit). Considering these two limiting cases for NH_4NO_3 , the amount of Ca^{2^+} plus K^+ in equivalence is 44% of NO_3^- for the data points with $\Delta NO_y < 2500$ pptv. Thus, when Ca^{2^+} and K^+ are taken into account, the lower limit of NH_4NO_3 was estimated to be 56% of the total NO_3^- .

[24] For $\Delta NO_y < 2500$ pptv, the correlation between Cl⁻ and Na⁺ showed a slope of 1.35 with $r^2 = 0.78$ (not shown), close to the seawater ratio of 1.16. Moreover, Mg^{2^+} showed a correlation with Na⁺ ($r^2 = 0.64$) with a slope of 0.13, very close to their seawater ratio of 0.11, suggesting that sea-salt particles partly contributed to these species. A Cl⁻/Na⁺ ratio lower than the seawater ratio indicates acidification of particles by HNO₃, resulting in a Cl⁻ deficit to form NaNO₃. By using these data points, NaNO₃ was estimated to contribute <12% of NO₃ for $\Delta NO_y < 2500$ pptv. Overall, more than 44% of particulate NO₃ ($\Delta NO_y < 2500$ pptv) was estimated to be in the form of NH₄NO₃.

[25] On the other hand, for $\Delta NO_y > 2500$ pptv, the ion balance between NO_3^- and $NH_4^+ - SO_4^{2-}$ (open circles in Figure 9) and the relative abundance of $Ca^{2+} + K^+$ compared to NO_3^- (Figure 8) suggest that more than 70% of the observed NO_3^- was in the form of NH_4NO_3 . In this case, the contribution of $NaNO_3$ was negligible (<1%).

[26] For the data points with $\Delta NO_y > 2500$ pptv (<3% of the total), however, NO_3^- (7500–12,000 pptv) far exceeded the 1:1 line relative to ΔNO_y (Figure 5b). As stated in section 2, there are two possible explanations for the excess NO_3^- , defined as δNO_3^- :

$$\delta NO_3^- = NO_3^- - \Delta NO_y \tag{4}$$

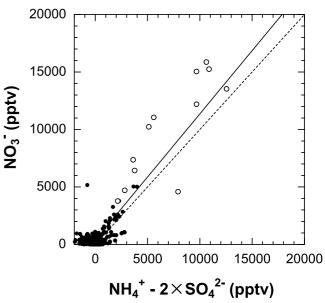


Figure 9. Relationship between NO_3^- and NH_4^+ minus SO_4^{2-} in equivalence. The solid line represents the least squares best fit, with a slope of 1.10. The dashed line indicates the 1:1 line. Open circles correspond to the data with high concentrations of Ca^{2+} (>400 pptv) obtained at 0–2 km over the Yellow Sea during flight 14.

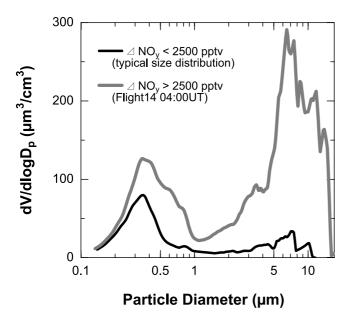


Figure 10. Typical size distributions representative of particles with $\Delta NO_y < 2500$ pptv (black line) and within the plumes with $\Delta NO_y > 2500$ pptv at 0400 UTC during flight 14 (gray line) measured by the OPC. The size distributions are calculated from number distributions for particles grown at ambient relative humidity, assuming that uptake is by dust particles.

(1) particles were not sampled by the NO_y instrument but were by the PILS-IC, because of their sizes, with $D_p = 1.5-3~\mu m$ in the different inlet systems, and (2) particles were measured as water-soluble ions by the PILS-IC but not converted to measurable NO in the NO_y instrument (i.e., $Ca(NO_3)_2$ and KNO_3). We explore these two possibilities in detecting particulate NO_3^- by the NO_y instrument and the PILS-IC system.

[27] Regarding particle size, the data points where $NO_3^ \Delta NO_{\nu}$ ratios exceed 1 (Figure 7), where coarse-particle volume tends to be larger relative to fine-particle volume under the high concentrations of Ca²⁺ (>400 pptv), are indicative of dust loading. Jordan et al. [2003b] used the DC-8 aircraft data for TRACE-P and showed that the dust was generally mixed with pollution and that a large fraction of nitrate was taken up by dust particles in the coarse mode, presumably through displacement of carbonate. Figure 10 shows OPC size distributions at ambient relative humidity with $\Delta NO_v > 2500$ pptv and typical distributions representative of particles for $\Delta NO_v < 2500$ pptv. The size distributions of particles with $\Delta NO_v < 2500$ pptv show that particles reside mostly in the fine mode ($D_p < 1 \mu m$), while the abundance in the coarse mode ($D_p > 1 \mu m$) is minimal. On the other hand, the distributions of particles with $\Delta NO_v > 2500$ pptv were characterized by the bimodal peaks within fine and coarse modes. On the basis of the size distributions and the PILS correction efficiency, we can roughly estimate that particles with $D_p = 1.5-3 \mu m$ could contribute to about 10% of the total volume of particles collected by the PILS for $\Delta NO_v > 2500$ pptv.

[28] We further examined particulate NO_3^- containing the 1.5–3- μ m range by using the $NO_{\nu}(f)$ data. Figure 11 shows

the relationship between NO_3^- and NO_3^- (L) (equation 3). Using the empirical expression given by Durham and Lundgren [1980], EF was estimated to be 2 for $D_p =$ 1.5 μ m, by assuming a particle density of 1.8 g cm⁻³ estimated from the OPC data. As stated in section 2, NO₃ (L) represents particulate NO₃ including larger particles $(D_p > 1.5 \mu m)$, mainly as NH₄NO₃, which can be detected by the NO_y instrument. Considering that the EF range is $2 \pm$ 0.5 for particles with $D_p = 0.5-3 \mu m$, the estimated uncertainty of NO_3^- (L) is given as +40%/-30%. For $\Delta NO_v < 2500$ pptv, NO_3^- showed agreement with NO_3^- (L), including particles that were not efficiently detected by the rearward NO_v system, indicating that almost all of δ NO₃ can be explained by NO_3^- (L). For $\Delta NO_v > 2500$ pptv, although it is difficult to evaluate the contribution of NO₃ (L) to δNO_3^- because of the uncertainty of the EF, we estimate that approximately $20 \pm 10\%$ of δNO_3^- can be explained by NO_3^- (L). The amount of NO_3^- (L) in this case corresponds to 10% of the PILS NO₃ in mass, which agrees well with the estimate from the particle volume by OPC (Figure 10), if the particle density is assumed to be constant.

[29] One of the possible reasons for the NH₄NO₃ particle size to reside in larger sizes ($D_p = 1.5 - 3~\mu m$) is that the relative humidity of the ambient air was relatively high, causing particles to absorb water before sampling. However, there is no clear relationship between the NO₃⁻/ Δ NO_y ratios and ambient relative humidity (not shown). Moreover, whether these particles were internally or externally mixed in the dust particles cannot be identified from these data alone.

[30] The sum of Ca^{2+} and K^+ in equivalence is plotted against δNO_3^- for $\Delta NO_y > 2500$ pptv in Figure 12. They are tightly correlated ($|\mathbf{r}|^2 = 0.93$), indicating that δNO_3^- is partly associated with dust particles. In the current analysis,

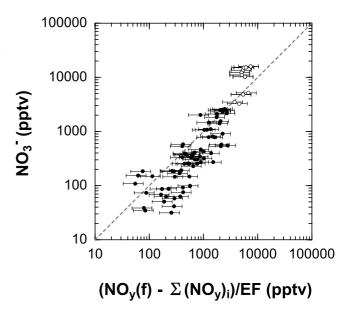


Figure 11. Scatterplot between NO₃ and NO₃ (L) = (NO_y) (f) $-\sum$ (NO_y)_i)/EF, where the EF (enhancement factor) is assumed to be 2. Open circles correspond to the data with Δ NO_y > 2500 pptv, while solid circles correspond to Δ NO_y < 2500 pptv. Error bars indicate the uncertainty in EF. The dashed line indicates the 1:1 line.

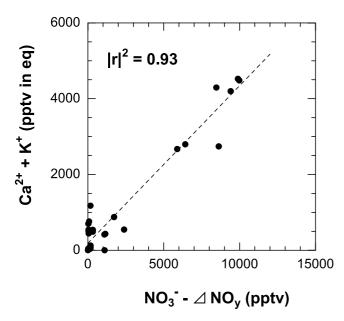


Figure 12. Scatterplot of the sum of Ca^{2+} and K^{+} in equivalence and δNO_{3}^{-} . The dashed line represents the least squares best fit.

 Ca^{2^+} and K^+ in equivalence accounted for 20% and 25% of δNO_3^- on average, respectively, for $\Delta NO_y > 2500$ pptv. Namely, $\text{Ca}(NO_3)_2$ and KNO_3 , which are not measured as NO_y , can explain a maximum of 45% of the δNO_3^- . The remaining δNO_3^- ($\sim\!35\%$), which cannot be explained by either larger NH_4NO_3 ($D_p=1.5-3~\mu\text{m}$) or Ca^{2^+} plus K^+ particles, is within the combined uncertainties of the measurements and EF.

5. Contribution of Particulate NO₃ to the NO_y Budget Over the Western Pacific

[31] Finally, we investigated the partitioning between particulate NO_3^- and gas-phase NO_{ν} . Figure 13 shows vertical profiles of NO_3^- , $\sum (NO_y)_i$, and $\sum (NO_y)_i + NO_3^-$ to NO_y ratios at $10^\circ - 30^\circ N$ and $30^\circ - 45^\circ N$. The partitioning of individual NO_v compounds is summarized in Table 3. The NO_v mixing ratios were systematically higher by 10– 30% than the sum of individual gas-phase NO_v compounds at 0-4 km. An important feature is that particulate NO₃ contributed 10-30% of the measured NO_{ν} at 0-4 km. On average, the particulate NO₃ fraction of NO_v was largest (26-37%) at 0-2 km, which was comparable to that of gasphase HNO_3 (24–32%) at the same altitude. The result indicates that the active conversion of HNO₃ to NO₃ is important, and particulate NO₃ may thus contribute a significant fraction (\sim 50%) to total nitrate, especially at 0-2 km during the TRACE-P period. Although the data for $\Delta NO_{\nu} > 2500$ pptv observed at 0-2 km are included in Figure 13, the fraction of these data was only <3% of the total. Therefore a majority of NO₃ shown in the figure are particles detected by the NO_v instrument.

[32] At altitudes of 2-6 km, very low values of NO_3^-/NO_y (0.02–0.10) indicate that a large fraction of the NO_3^- was scavenged by cloud droplets during the vertical transport from lower altitudes (0–2 km) [Koike et al., 2003;

Kondo et al., 2004]. Note that the NO_3^- concentrations discussed here are fine-mode fractions of the total particulate NO_3^- , which were obtained by the PILS having a 50% size cut of 1.3 μm. When we use the NO_3^- data obtained by the DC-8 filter samples, which measured particles with D_p up to 6 μm [*Dibb et al.*, 2002], particulate NO_3^- constituted 54% of the total nitrate on average, reaching a maximum of 72% at 0–2 km where observed particles were heavily influenced by dust during TRACE-P [*Jordan et al.*, 2003b].

[33] For gas-phase NO_{ν} , 80–90% of the measured NO_{ν} at 0-6 km is accounted for by NO_x (= $NO + NO_2$), PAN, and HNO₃. These values are similar to those in air masses influenced by continental surface emissions previously observed at 0−5 km over the western Pacific during the Pacific Exploratory Mission (PEM) West B in 1994 [Kondo et al., 1997b]. It is noted that gas-phase HNO₃ was measured using a chemical ionization mass spectrometer (CIMS) onboard the P-3B during TRACE-P. During PEM-West B and TRACE-P, on the other hand, HNO3 was measured by the University of New Hampshire's (UNH) mist chamber/ion chromatography (MC-IC) instrument onboard the DC-8 research aircraft. The UNH/MC-IC could sample nitrate-containing particles, as indicated by intercomparison between the two instruments during TRACE-P [Zondlo et al., 2003]. Accordingly, the sampling by the UNH/MC-IC would result in higher measurements of gasphase HNO₃ relative to the true amount. Thus much more accurate evaluations of the NO_v budget are enabled in this study by highly selective and sensitive measurements of gas-phase HNO₃ by the CIMS instrument and of particulate NO_3^- by the PILS-IC.

[34] PAN was systematically higher at $30^{\circ}-45^{\circ}N$ (175–930 pptv) than at $10^{\circ}-30^{\circ}N$ (140–190 pptv), consistent with its chemical stability at lower temperatures. Meanwhile, PPN (<6%) and alkyl nitrate (<5%) were minor compounds in the NO_v budget during TRACE-P. The

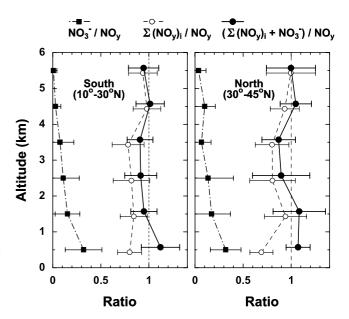


Figure 13. Vertical profiles of the mean values of NO_3^-/NO_y (solid squares), $\sum (NO_y)_i/NO_y$ (open circles), and $(\sum (NO_y)_i + NO_3^-)/NO_y$ (solid circles) at (left) $10^\circ - 30^\circ N$ and (right) $30^\circ - 45^\circ N$.

Table 3. Reactive Nitrogen Budgets and Partitioning at 10°-30°N and 30°-45°N on Board P-3B During TRACE-P^a

Altitude, km	Latitude ^b	NO ₃ , pptv	NO_x , pptv	HNO ₃ , pptv	PAN, pptv	PPN, pptv	Alkyl Nitrate, pptv	NO _y , pptv	$\Sigma(NO_{\rm v})_i + NO_3^-, pptv$
4-6	N	25 ± 42	34 ± 37	115 ± 81	174 ± 122	21 ± 19	15 ± 7	388 ± 238	359 ± 203
		(0.06 ± 0.12)	(0.08 ± 0.08)	(0.34 ± 0.20)	(0.47 ± 0.11)	(0.05 ± 0.02)	(0.05 ± 0.02)	(-)	(1.05 ± 0.23)
4-6	S	18 ± 18	30 ± 15	156 ± 70	143 ± 104	14 ± 10	9 ± 4	368 ± 174	437 ± 229
		(0.02 ± 0.04)	(0.08 ± 0.03)	(0.44 ± 0.16)	(0.35 ± 0.13)	(0.03 ± 0.01)	(0.03 ± 0.01)	(-)	(0.98 ± 0.16)
$^{2-4}$	N	103 ± 408	70 ± 91	221 ± 139	438 ± 258	61 ± 38	30 ± 11	997 ± 628	883 ± 747
		(0.10 ± 0.21)	(0.07 ± 0.05)	(0.23 ± 0.14)	(0.46 ± 0.13)	(0.06 ± 0.02)	(0.04 ± 0.01)	(-)	(0.91 ± 0.29)
$^{2-4}$	S	110 ± 237	63 ± 47	289 ± 308	160 ± 142	24 ± 13	15 ± 7	713 ± 655	889 ± 628
		(0.09 ± 0.16)	(0.10 ± 0.05)	(0.39 ± 0.17)	(0.21 ± 0.14)	(0.03 ± 0.02)	(0.04 ± 0.03)	(-)	(0.91 ± 0.15)
0 - 2	N	1761 ± 3199	1142 ± 4948	430 ± 316	928 ± 542	155 ± 91	63 ± 24	3836 ± 5865	3943 ± 4749
		(0.37 ± 0.39)	(0.16 ± 0.15)	(0.24 ± 0.20)	(0.33 ± 0.10)	(0.05 ± 0.02)	(0.03 ± 0.01)	(-)	(1.18 ± 0.34)
0 - 2	S	230 ± 432	146 ± 200	209 ± 222	189 ± 184	47 ± 34	34 ± 18	776 ± 624	1011 ± 556
		(0.26 ± 0.19)	(0.16 ± 0.10)	(0.32 ± 0.17)	(0.20 ± 0.12)	(0.05 ± 0.03)	(0.05 ± 0.03)	(-)	(1.06 ± 0.19)

^aValues are mean $\pm 1~\sigma$. The numbers in parentheses are the ratios to the measured NO₃. The numbers of data points are within a range of 40-120 in each column.

present results emphasize that the contribution of particulate NO_3^- to NO_y should be taken into account when evaluating NO_y speciation and budget closure using an NO_y instrument with a catalytic converter, especially when near a large source region of particulate NO_3^- .

6. Conclusions

[35] The sum of individual NO_y gas components $(\sum (NO_y)_i)$, NO_y, and particulate NO₃ (D_p <1.3 µm) were measured simultaneously on board the P-3B aircraft at 10° – 45°N over the western Pacific during TRACE-P. The inorganic chemical composition of aerosols, including NO₃, was measured by the PILS-IC technique. In addition to the field measurements, conversion efficiencies of particulate NH₄NO₃, KNO₃, NaNO₃, and Ca(NO₃)₂ by the gold tube catalytic converter were measured in the laboratory. NH₄NO₃ was converted to NO with an efficiency of 91 ± 9%, which was as high as that for HNO₃. Meanwhile, the conversion efficiencies of the aerosols in the other forms were as low as 1–3%. These results are consistent with the particulate NO₃ measurements made on board the aircraft.

[36] Elevated NO₃ was typically observed within biomass/biofuel burning plumes at 0–4 km. The measured NO_y was systematically higher by 10–30% than the $\sum (\text{NO}_y)_i$ at 0–4 km. Under low dust loadings with $[\text{Ca}^{2+}] < 200$ pptv, the observed particles reside mostly in the fine mode, while the size distributions showed bimodal distribution within the fine and coarse modes under high dust loadings ($[\text{Ca}^{2+}] > 400$ pptv). The $\Delta \text{NO}_y = \text{NO}_y - \sum (\text{NO}_y)_i$ values were approximately equal to NO₃ concentrations under low dust loadings, suggesting that the NO_y instrument quantitatively converted NO₃ to NO. More than 44% of the NO₃ detected by the NO_y instrument was estimated to be in the form of NH₄NO₃, which was identified from the measured ionic composition of the aerosols.

[37] NO_3^- far exceeded ΔNO_y only in polluted air masses with high dust loadings at 0-2 km over the Yellow Sea $(30^\circ-45^\circ N)$. By using the forward facing NO_y inlet, 20% of the excess $NO_3^ (\delta NO_3^- = NO_3^- - \Delta NO_y)$ was estimated to be mainly in the form of NH_4NO_3 , suggesting that the size of this NO_3^- particle species detected by PILS-IC was in the range of $D_p = 1.5-3$ μm , which was larger than the size cutoff of the rearward facing NO_y instrument. Some of the

other δNO_3^- can be in the form of $Ca(NO_3)_2$ (<20%) and KNO_3 (<25%) under high dust loading with a mixture of anthropogenic pollution, considering the measured aerosol ion composition. Particulate NO_3^- in these forms can partly explain the large δNO_3^- , because their conversion in the NO_ν instrument is negligible. The remaining δNO_3^- (<35%), which cannot be explained either by larger NH_4NO_3 ($D_p=1.5-3~\mu m$) or $Ca^{2+}+K^+$ particles, is within the combined uncertainties of the measurements and the enhancement factor for the forward facing NO_ν inlet.

[38] Generally, the contribution of particulate NO_3^- to NO_y was most important at 0-2 km, where the NO_3^- concentrations constituted 10-30% of NO_y . On average, the concentrations of NO_3^- and gas-phase HNO_3 were comparable in this region. These results indicate the importance of properly including particulate NO_3^- in assessing the budget of reactive nitrogen using measurements of NO_y by the catalytic conversion technique, especially when near a large source region of particulate NO_3^- .

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^bLatitudes 10°-30°N are denoted as S, and latitudes 30°-45°N are denoted as N.

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